# A guide to monitoring the ecological quality of ponds and canals using PSYM 

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# MONITORING THE QUALITY OF STILL WATERS USING PSYM 

## 1. Introduction

PSYM, the Predictive SYstem for Multimetrics, (pronounced sim) has been developed to provide a method for assessing the biological quality of still waters in England and Wales.

The method uses a number of aquatic plant and invertebrate measures (known as metrics) ${ }^{1}$, which are combined together to give a single value which represents the waterbody's overall quality status.

Using the method involves the following steps:

1. Simple environmental data are gathered for each waterbody from map or field evidence (area, grid reference, geology etc.).
2. Biological surveys of the plant and animal communities are undertaken and net samples are processed.
3. The biological and environmental data are entered into the PSYM computer programme which:
(i) uses the environmental data to predict which plants and animals should be present in the waterbody if it is undegraded,
(ii) takes the real plant and animal lists and calculates a number of metrics ${ }^{1}$.

Finally the programme compares the predicted plant and animal metrics with the real survey metrics to see how similar they are (i.e. how near the waterbody currently is to its ideal/undegraded state). The metric scores are then combined to provide a single value which summarises the overall ecological quality of the waterbody. Where appropriate, individual metric scores can also be examined to help diagnose the causes of any observed degradation (e.g. eutrophication, metal contamination).

## 2. Background

### 2.1 Why was the method developed?

Historically, the Environment Agency and other statutory bodies have undertaken relatively little monitoring of still waters (lakes, ponds, canals, ditches etc.). The absence of a standardised assessment method was a major barrier to the assessment of these waterbodies.
The PSYM methodology provides a standard assessment method for still waters which enables a variety of organisations involved in waterbody management to consider water quality in a broad national context. It provides the Environment Agency with a means to assess still water quality for General Quality Assessment (GQA) and other reporting purposes, and can be used in partnership with others such as DEFRA or English Nature. The method also enables public or private sector NGOs (e.g. consultants, community groups) to improve general standards of assessment in waterbody management plans or environmental impact assessments, and provides a means of assessing management techniques.

### 2.2 About PSYM

PSYM is a waterbody quality assessment methodology which essentially combines the predictive approach of RIVPACS ${ }^{2}$ with multimetric-based methods used for ecological quality assessment in the United States.

In multimetric assessments, a range of variables (metrics) each related to degradation is used to assess water quality giving a broad-based assessment of quality. The values from individual metrics are combined to give a single measure which aims to represent the overall ecological quality of the waterbody. Combining this with predictive techniques gives a powerful method for comparing waterbodies of any type with their undegraded counterpart.

The PSYM methodology directly parallels the approach defined in the EU Water Framework Directive. This includes requirements for (i) comparisons with minimally impacted baseline conditions, and for (ii) assessments to be based on multiple parameters related to degradation.

### 2.3 Which waterbodies can be monitored using the method?

The PSYM approach is potentially applicable to all still waterbody types (e.g. lakes, ponds, temporary ponds, canals). However, to apply the method, specific data need to be collected from each waterbody type. These data are used both to (i) develop equations which can be used to predict the species which should occur at an undegraded site and (ii) to identify which biotic measures (e.g. species richness, ASPT) are the most effective at tracking degradation in that waterbody type.

[^0]So far, the method has been developed for use on two still waterbody types (i) canals (ii) ponds ${ }^{3}$ and small lakes (up to about 5 ha in area). An extension of the method for temporary ponds is currently being developed independently by PCTPR with support from the Freshwater Biological Association. Methods have not, so far, been developed for assessing the quality of large lakes, ditches or brackish waters.

The baseline dataset used to develop the metrics for ponds was based on survey data from sites with broad coverage of England and Wales from a wide range of altitudes ( $0-550 \mathrm{~m}$ ), and land types (representative coverage of ITE land classes), so the resulting model is suitable for sites across England and Wales.

### 2.4 Why assess water quality using both plants and invertebrates?

Ideally, PSYM should use information from both the plant and animal communities present in a waterbody. This is because, together, plants and animal groups span a complementary range of sensitivities to potential degradation factors. Plants are, for example, particularly sensitive to waterbody nutrient status, whereas animals typically exhibit greater oxygen sensitivity.

Matrix analysis suggests that in most waterbodies, the most effective plant group to use for assessment is likely to be either diatoms or macrophytes. The most effective animal groups are likely to be macroinvertebrates and/or potentially fish in large permanent waters. Combining a plant and animal group from these assemblages gives a range of taxa which span a number of trophic levels, occupy a variety of waterbody habitats (e.g. can be found in the littoral zone and open water) and are long-lived, so that they can provide a temporally and spatially integrated measure of the current ecosystem state. Invertebrate, diatom and macrophyte assemblages are also relatively species-rich groups, ensuring that a good cross section of waterbody biodiversity is included in the quality assessment.

In ponds, macroinvertebrates and macrophytes have been chosen as the most practical and effective taxa for quality assessment. In canals, the choice was macroinvertebrates and diatoms, although the method has so far only been developed for macroinvertebrates. Macrophytes were assessed as being less suitable for canal assessment because the high turbidity and artificial banks which characterise most navigated canals often means that very few higher plant species are present, regardless of overall water quality.

### 2.5 Do you have to use both plant and invertebrates for PSYM pond assessments?

Although PSYM pond quality assessments should be made using both plant and invertebrate assemblages, a partial assessment can be made using just one assemblage if necessary. If this is the case, macroinvertebrates are likely to be the best single choice of organisms for assessing overall waterbody quality. Macrophytes, however, have the advantage of being very quick to survey and can be used, if necessary, as a rapid bio-assessment method.

### 2.6 How are the plant and invertebrate metrics chosen?

Metrics are biological measures (such as taxa richness) which vary with anthropogenic degradation and can, therefore, be used to measure the extent of ecosystem degradation. The concept underlying multimetric assessment is that by using a number of different measures and summing these together, an overall assessment of environmental degradation can be made. For canals, at present, only an invertebrate option is available.

Metrics are chosen by correlating known degradation gradients (nutrient levels, heavy metal levels, presence of road runoff etc.) with a wide list of possible test metrics e.g. family richness, number of exotic species, EPT (number of Ephemeroptera, Plecoptera and Trichoptera families). The 'test' list is narrowed-down to a list of viable metrics by looking at the significance of relationships between each potential metric and anthropogenic degradation gradients. For invertebrates, metrics are chosen at the highest taxonomic level i.e. family or order level rather than species-level to reduce effort (although species level information can be derived from the samples if needed for conservation work). In practice, there were generally at least equally strong correlations between family-level macroinvertebrate metrics and degradation as there were between species-level metrics and degradation. This enables family-level macroinvertebrate data to be used for quality assessments in both ponds and canals. Plant metrics are generally based on species level information.

[^1]Analyses have shown that the most effective metrics for assessing environmental degradation in ponds and canals are:

## Ponds

## Invertebrates

- Average score per taxon (ASPT)
- Number of dragonfly (Odonata) and alderfly (Megaloptera) families (F_OM)
- Number of beetle (Coleoptera) families (F_COL)


## Plants:

- Number of submerged and emergent plant species (SM_NTX)
- Trophic ranking score for aquatic and emergent plants (TRS_ALL)
- Number of uncommon plant species ((PL_NUS)


## Canals

## Invertebrates

- Average score per taxon (ASPT)
- Number of Ephemeroptera, Plecoptera and Trichoptera families (F_EPT)
- Number of beetle families (F_COL)
- Number of invertebrate families (INV_NFA)

Note that in canals methods for assessing the chosen plant group (diatoms) have not yet been developed.

In order to calculate predictions for these metrics the PSYM model predicts which taxa will be found at a site. An example of a predicted and observed taxa list is given in the following table.

Predicted and observed taxa lists for pond plants and macroinvertebrates for Asham Meads field pond, Oxfordshire.

| Species | Predicted (probability of occurrence) | Observed | Species | Predicted (probability of occurrence) | Observed |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Wetland plants |  |  | Macroinvertebrates |  |  |
| Agrostis stolonifera | 0.76 | $\checkmark$ | Lymnaeidae | 1.00 | $\checkmark$ |
| Juncus effusus | 0.75 | $\checkmark$ | Planorbidae | 1.00 | $\checkmark$ |
| Epilobium hirsutum | 0.66 | $\checkmark$ | Glossiphoniidae | 1.00 | $\checkmark$ |
| Solanum dulcamara | 0.64 | $\checkmark$ | Coenagrionidae | 1.00 |  |
| Juncus articulatus | 0.61 | $\checkmark$ | Corixidae | 1.00 | $\checkmark$ |
| Alisma plantagoaquatica | 0.58 | $\checkmark$ | Haliplidae | 1.00 | $\checkmark$ |
| Glyceria fluitans | 0.54 | $\checkmark$ | Dytiscidae | 1.00 | $\checkmark$ |
| Typha latifolia | 0.52 |  | Hydrophilidae | 1.00 | $\checkmark$ |
| Lycopus europaeus | 0.52 |  | Notonectidae | 0.80 | $\checkmark$ |
| Mentha aquatica | 0.50 | $\checkmark$ | Baetidae | 0.78 | $\checkmark$ |
| Juncus inflexus | 0.48 | $\checkmark$ | Asellidae | 0.76 | $\checkmark$ |
| Galium palustre | 0.43 | $\checkmark$ | Libellulidae | 0.75 |  |
| Sparganium erectum | 0.42 |  | Gerridae | 0.64 | $\checkmark$ |
| Eloeocharis palustris | 0.39 | $\checkmark$ | Leptoceridae | 0.61 |  |
| Deschampsia caespitosa | 0.38 | $\checkmark$ | Sialidae | 0.61 |  |
| Myosotis scorpioides | 0.30 | $\checkmark$ | Hydraenidae | 0.58 | $\checkmark$ |
|  |  |  | Limnephilidae | 0.56 | $\checkmark$ |
| Aquatic plants |  |  | Aeshnidae | 0.53 |  |
| Lemna minor | 0.67 | $\checkmark$ | Crangonyctidae | 0.49 | $\checkmark$ |
| Callitriche spp. | 0.52 | $\checkmark$ | Caenidae | 0.45 | $\checkmark$ |
| Chara spp. | 0.44 |  | Planariidae | 0.42 |  |
| Potamogeton natans | 0.32 | $\checkmark$ | Erpobdellidae | 0.39 |  |
|  |  |  | Hydrobiidae | 0.32 |  |

### 2.7 Which physical and chemical variables are used in the predictions?

As in RIVPACS, the PSYM method assesses quality by comparing actual and predicted quality scores for each waterbody. The predictions of unimpaired waterbody quality are made using physico-chemical data gathered from the waterbody.

In ponds the main predictors of unimpaired community type fall into nine major variable categories. Of these, three are relatively invariant (e.g. grid reference, altitude, base geology) which need only be assessed once. The remaining six categories of variables require on-site field measurement when each assessment is made. These are area, pH , shade, grazing, presence of an inflow and emergent plant cover. In canals, the main predictive variables are grid reference, altitude, alkalinity, substrate and boat traffic.

### 2.8 How are metrics scored?

When a waterbody is assessed, each individual metric is calculated and compared to the computer predicted score for that metric. The relationship between observed and expected is presented as a percentage of similarity, and then transformed to a 4 point scale e.g. $0,1,2$ and 3 where 0 represents poor quality, and 3 represents good quality (i.e. no deviation from expected). All metric scores are then summed to give an overall quality index, which is presented as a percentage of the maximum score and, potentially, forms the basis of General Quality Assessment (GQA) categorisation of a site.

### 2.9 Diagnosis

The main objective of the PSYM method is to assess the overall condition of freshwater ecosystems. The system does not, in itself, diagnose the cause, or causes, of degradation. Indeed it is considered inappropriate for a general quality assessment method to be biased towards the evaluation of a single impact. However, there is considerable potential for data which are collected using the scheme to be re-interpreted to diagnose the causes of degradation. Individual metrics can indicate aspects of water quality and the raw data can be reanalysed to give pollution indices, such as trophic scores or acidification indices.

## 3. Assessing pond quality using Pond PSYM

### 3.1 Introduction

Pond PSYM has currently been developed for use in the Summer season (June, July, August), and is based on assessments of both macroinvertebrate and macrophyte assemblages.

### 3.2 Sites which can be included

Pond PSYM can be used on ponds and small lakes up to about 5 ha in area in England and Wales. The method can, in theory be used to assess the quality of seasonal ponds, but in practice it 'over-predicts' for ponds which are highly seasonal (i.e. which dry hard every year), and is best restricted to ponds which are either permanent, or semi-seasonal (i.e. which dry occasionally in very hot years). An extension of the method is currently being developed for use with fully temporary ponds.

### 3.3 Field data collection

The environmental data which need to be collected from each pond to use Pond PSYM include:
(i) locational and other data used for data processing. This includes: site name and code, county and nearest town, six or eight figure grid reference as necessary to identify the site, survey date, surveyor, site description.
(ii) predictive variables used in the pond PSYM programme to predict the undegraded biota for the pond. This includes: map-based locational information (six figure grid reference, altitude), together with site data describing shade, the presence of an inflow, cover of emergent plants, pond base geology and pH .

## Collecting predictive variable data

The methods used to collect the main predictive variable data are briefly outlined below.
Grid reference: six figure reference, taken from 1:50,000 or 1:25,000 OS maps, input into the model as Easting and Northing ( 100 km cell reference followed by 3 figures).

Altitude: in metres above sea level, taken from 1:50,000 or 1:25,000 OS maps.
$\boldsymbol{p} \boldsymbol{H}$ : measured either (i) in the field in a bucket of water taken from a representative area of the pond, or (ii) using a water sample collected in the field and analysed later in the laboratory. For laboratory analysed samples, use acid washed bottles stored in a cool place after collection (e.g. cold box) and analyse within one day of collection.

Pond area: this is the area lying within the outer edge of the pond (see 3.4 below). The pond dimensions can be measured using a tape, or by careful pacing. A small sketch can help to make this estimate. For large ponds it can be easier to use an OS map outline, with the dimensions checked in the field. Note that for the predictions, area data are used as $\log$ values so, particularly for large ponds, estimates do not need to be highly accurate.

Pond overhung: the percentage of the pond area which is directly overhung (e.g. by trees, scrub etc.).
\% of pond edge grazed by livestock: the percentage of the perimeter of the pond to which livestock have active access. Note that if cattle, sheep, horses etc. are not grazing at the time of the survey, their presence can be detected by other features such as poaching of the ground.

Pond base: the rock type underlying the pond (beneath the sediment). This can often be assessed directly in the field, or be determined using a geology map. In the field, push the handle of the pond net through the sediment into the base. Exact measurement is not necessary, only broad categorization into one of three percentage categories: $1=0 \%-32 \%$, $2=33 \%-66 \%, 3=67 \%-100 \%$.

Inflow: whether or not the pond has a surface inflow. This can be a direct or indirect inflow from a river, stream, ditch, spring or seepage. The inflow can be $d r y$ at the time of the survey.

Emergent plant cover: the percentage of the pond covered by emergent plant species. The term 'emergent plant species' includes all species listed as emergents on the wetland plant recording sheet. It includes these species regardless of their habit at the time of the survey (e.g. some emergent species may be growing predominantly under water at the time of the survey). It does not include any other species e.g. terrestrial species or plants specifically defined as 'submerged' or 'floating-leaved' plant species on the wetland plant recording sheet.

Estimates of the percentage cover of emergent plants should be made for the whole area within the outer edge of the pond, not the current water area. The cover of sparsely growing stands of plants (e.g. occasional bulrush plants with much open water between), should be estimated as if they were growing closely together. The easiest way of doing this is to imagine all emergent plants pushed together on one side of the pond, with an estimate then made of what proportion of the pond this covers.

At present it is recommended that for those variables for which field estimates are made ( pH , area, overhanging trees, grazing, base type and emergent plant cover) the objective of measurement should be to obtain estimates that are within
$5-10 \%$ of the long term mean. It is expected that further work will be undertaken to refine understanding of the effects of variation in measurements in the future.

### 3.4 Defining the outer edge of the pond

Identifying the 'outer edge' of the pond is important for many of the physico-chemical survey assessments and for undertaking the plant survey. In all cases, the definition of pond 'outer edge' is 'the upper level at which water stands in winter'.

In practice, the outer edge is usually readily discernible from one or more site characteristics. The best of these is usually the distribution and/or morphology of wetland plants. For example, it may be marked by a fringe of soft rush (Juncus effusus) or by thick bundles of fine roots growing out of the trunks of willows etc. Alternatively, the line can often be seen as a 'water mark' on surrounding trees or walls and is sometimes evident as a break of slope. The outer boundary of the pond will usually, of course, be dry at the time of the survey.

### 3.5 Plant survey methodology

The aim of plant recording is to make a complete list of wetland plants present within the outer edge of the pond. The field recording sheet gives a definitive list of the plant species regarded as 'wetland'. Terrestrial plants and wetland plants growing outside the outer edge of the pond are not recorded. The wetland plant recording sheet includes submerged macrophytes, floating-leaved species and emergent macrophytes, and these groups are used separately in analysis.

Pond macrophytes are surveyed by walking or wading the entire perimeter of the dry and shallow water areas of the waterbody. Deeper water areas are sampled either using a pond net or by grapnel thrown from shallow water or from a boat.

Most wetland plants are readily identifiable using a hand lens. However, with a few species (especially fine-leaved Potamogeton and Callitriche spp.) it may be necessary to remove a small amount of plant material for later microscopic examination and confirmation.

Record macrophyte species found on the attached wetland plant recording sheet.

### 3.6 Invertebrate survey methodology

The pond invertebrate survey methods used for PSYM are based on standard three minute hand-net sampling methods developed for the National Pond Survey (Pond Action, 1998).

The NPS invertebrate survey techniques were developed 'post-RIVPACS' in 1989-90, and were designed to be closely compatible with the original RIVPACS sampling methods, whilst allowing for differences between river and pond habitat types. The main differences between pond and river sampling methods are that:

- RIVPACS allocates sampling time on an area basis (i.e. more time is spent sampling extensive habitats). In pond PSYM, time is allocated according to mesohabitat types (i.e. if six main habitat types are identified time is divided equally amongst these). This change was made to allow for the fact that many ponds have extensive biologically uniform areas of open water and silt, and narrow but highly diverse marginal zones.
- In Pond PSYM the 3 minute survey subsamples are taken around the entire pond site whereas in RIVPACS samples are collected from an area that can be covered comfortably in three minutes: typically a river length of 5-20 m.


### 3.7 Selecting mesohabitats for invertebrate surveys

All the main mesohabitats in the pond are sampled so that as many invertebrate species are collected from the site as possible. Examples of typical mesohabitats are: stands of Carex (sedge); gravel- or muddy-bottomed shallows; areas overhung by willows, including water-bound tree-roots; stands of Elodea, or other submerged aquatics; flooded marginal grasses; and inflow areas. As a rough guide, the average pond might contain 3-8 mesohabitats, depending on its size and complexity. It is important that vegetation structure, as well as plant species composition, is considered when selecting mesohabitats: it is better to identify habitats consisting of e.g. soft floating leaves, stiff emergent stems, etc. than to make each different plant species a separate habitat. Mesohabitats are identified during the initial walk around the pond examining vegetation stands and other relevant features (this can be combined with the initial plant survey stage).

## Invertebrate sampling method

(i) The three-minute sampling time is divided equally between the number of mesohabitats recorded: e.g. for six mesohabitats, each will be sampled for 30 seconds. Where a mesohabitat is extensive or covers several widelyseparated areas of the pond, the sampling time allotted to that mesohabitat is further divided in order to represent it adequately (e.g. into $6 \times 5$ second sub-samples).
(ii) Each mesohabitat is netted vigorously to collect macroinvertebrates. Stony or sandy substrates are lightly 'kicksampled' to disturb and capture macroinvertebrate inhabitants. N.B. deep accumulations of soft sediment are avoided, since these areas typically support few species and collecting large amounts of mud makes later sorting extremely difficult. Similarly, large accumulations of plant material, root masses, and the like should not be taken away in the sample: the idea is to dislodge and capture the animals without collecting an unmanageable sample.

The sample is placed in the labelled bucket for later sorting in the laboratory. Note: the three-minute sampling time refers solely to 'net-in-the-water' time, and does not include time moving between adjacent netting areas around the pond.
(iii) Amphibians or fish caught whilst sampling are noted on the recording sheet and returned to the pond.

## Additional invertebrate sampling

A further 1 minute (total time, not net-in-the-water time) is spent searching for animals which may otherwise be missed in the 3 -minute sample. Areas which might be searched include the water surface (for whirligig beetles, pond skaters etc.) and under stones and logs (for limpets, snails, leeches, flatworms etc.). Additional species found are added to the main 3-minute sample.

### 3.8 Processing invertebrate samples

Invertebrate sorting and identification methods follow the standard laboratory techniques. Invertebrate samples are identified to family level for most groups and class level for oligochaetes.

Record findings in the columns on the field sheet as follows. If present and so included in ASPT calculation, record in the "ASPT" column, if a dragonfly or alderfly family also record in the "OM" column, or if a Coleoptera family in the "Cole." column.

### 3.9 Data processing and analysis

Biotic data are used by pond PSYM to calculate three plant metrics and three invertebrate metrics:

## Plants:

- Number of submerged and emergent plant species (PL_NTX)
- Trophic ranking score for aquatic and emergent plants (TRS_ALL)
- Number of uncommon plant species (PL_NUS)


## Invertebrates:

- Average score per taxon (ASPT)
- Number of dragonfly (Odonata) and alderfly (Megaloptera ${ }^{4}$ ) families (F_OM)
- Number of beetle (Coleoptera) families (F_COL).


## Calculating the pond metrics from taxon lists

## 1. Number of submerged and emergent plant species

This is simply the sum of the number of submerged plant taxa plus number of emergent plant taxa observed at the site. The terms 'submerged' and 'emergent' taxa refer only to the species listed in these groups on the field sheet - not to plants of any species which happen to be submerged below water or growing round the edge of the pond at the time of the survey.

The calculation does not include the number of floating-leaved species present. This is because the pond data suggest that the number of floating-leaved plants occurring at a site does not decline significantly with increasing degradation. The metric is therefore improved by omitting this plant group.

[^2]
## 2. Trophic Ranking Score (TRS)

TRS is a measure of the average trophic rank for the pond. This is calculated by assigning each plant species with a trophic score based on its affinity to waters of a particular nutrient status. The trophic scores used in the present study were based on work undertaken on lakes by Palmer (1989). Plant scores in this system vary between 2.5 (dystrophic i.e. very nutrient poor conditions) and 10 (eutrophic, i.e. nutrient rich conditions).

Unfortunately, not all plants have trophic scores. This situation has arisen because the current TRS values for standing waters (Palmer et al., 1992) are based only on analysis of lake data, and many plant species which are common in ponds occurred at too low a frequency in lakes to give them a score. Nigel Holmes's Mean Trophic Ranking method, which was developed for assessing the nutrient status of running water communities, cannot be used in the current analysis because trophic values for some plant species can vary between still and running waters ( N . Holmes pers. comm.).

The TRS value for a site is calculated as follows:
(i) The trophic scores from each plant species present at the site are summed together.
(ii) The summed score is divided by the total number of plant species which have a trophic ranking score (NOTE not the total number of plants at the site) to give the TRS.

## 3. Uncommon species index

Uncommon species are those which have a rarity score of 2 or more. The number of these species is simply summed to give the number of uncommon species.

Uncommon species refers to species which can be described as 'local', 'nationally scarce' or 'Red Data Book'.
Descriptions of these categories are given below.

| Status | Rarity score | Definition |
| :--- | :---: | :--- |
| Common | 1 | Recorded from $>70010 \times 10 \mathrm{~km}$ grid squares in Britain |
| Local | 2 | Recorded from between 101 and 700 grid squares in Britain |
| Nationally Scarce | 4 | Nationally Scarce. Recorded from 15-100 grid squares in Britain |
| At risk | 8 | Red Data Book: Category "At risk" |
| Vulnerable | 16 | Red Data Book: Category "Vulnerable" |
| Endangered | 32 | Red Data Book: Categories "Endangered" or "Highly Endangered" |

The rarity score for each species is given on the plant recording sheet so the number of species with a rarity score of 2 or more can be easily calculated.

## 4. Average Score Per Taxon (ASPT)

ASPT is calculated, as in RIVPACS, by summing the BMWP ${ }^{6}$ scores for all taxa present at the site and dividing by the total number of BMWP taxa present.

## 5. Number of dragonfly and alderfly families

This metric is the sum of the number of Odonata and Megaloptera families which occur at the site.

## 6. Number of beetle families

This metric is the sum of the number of Coleoptera families present at the site. The metric has a relationship with bank quality as well as water quality.

[^3]
## 4. Assessing canal quality using Canal PSYM

### 4.1 Introduction

Canal PSYM has currently been developed for use in the Spring season (March, April, May), and is based on a macroinvertebrate assessment only ${ }^{7}$. Two canal PSYM models have been developed in response to the potential problem of obtaining bottom samples. The basic model uses combined edge and bottom samples, but where this is not possible, a second model can be used for which only edge samples are taken.

### 4.2 Sites which can be included

Canal PSYM can be used to assess the quality of any section of canal, including both reinforced and natural bank sections. The term canal, does not however include major navigations (i.e. canalised rivers), such as the Lee Navigation and Stort Navigation, since these were excluded from the canal survey as many sections are essentially riverine in character.

### 4.3 Field sheet data collection

Field data collected from each canal site include:
(i) locational and other data used simply to identify the site and enable the site to be re-found for monitoring purposes. These data include information on: site name and collection code, canal name, nearest town, six or eight figure grid reference (depending on the degree of accuracy needed to locate the site precisely), survey date, surveyor, description of site.
(ii) predictive variables used in the PSYM programme to predict the minimally impaired biota for the canal. This includes map- or desk-based information (grid reference, altitude, number of boats) and field-based measurements (alkalinity, canal substrate).

## Field variables

The environmental data which need to be collected from each site to use Canal PSYM depend on whether (i) only edge samples are taken or (ii) combined edge and bottom samples are used. For (i) Northing, altitude, turbidity, substrate and boat traffic are required. For (ii) Easting, Northing, altitude, alkalinity, substrate and boat traffic are needed. Details are as follows.

Easting: 100 km cell reference followed by 4 figures, from 1:25,000 OS maps.
Northing: 100 km cell reference followed by 4 figures, from 1:25,000 OS maps.
Altitude: in metres above sea level, taken from 1:50,000 or 1:25,000 OS maps.
Turbidity: Secchi depth in cm.
Total Alkalinity: measured as meq $\mathrm{l}^{-1}$. Analysed in the laboratory from a water sample collected in the field.
Canal substrate: a field estimate of the percentage of the canal sediment composition that is sand. Sediment composition often varies across the canal, with the edge area usually coarser than the bottom substrate in deeper water. Where this is the case, two substrate measurements should be made, one in shallow water and one in deep water and the average calculated.

Number of boats: measured in thousands of boat movements per annum. These data can be provided by British Waterways (or other canal authority as appropriate).

### 4.4 Invertebrate sampling

Canals are steep-sided and relatively deep waterbodies, so the area-related hand-net sampling methodologies appropriate for rivers (e.g. typical RIVPACS sampling) cannot be directly applied to canals. In particular: (i) hand-net methods are difficult to apply to the deepest open-water areas of canals, (ii) most invertebrate species are concentrated in a narrow band at the canal edge, so that an area-based sampling method can considerably under-sample invertebrate diversity.

The sampling technique used to collect invertebrate samples for this was developed as a hybrid between the 'threeminute hand-net sample' currently used for sampling shallow rivers, and the 'one-minute hand-net sample + dredge hauls' method recommended for sampling deep rivers. The method will also be used by CEH in future canal surveys.

[^4]The method comprises:

1. A one-minute search.
2. A two-minute semi-continuous hand-net sampling of the canal margin, shallows and any emergent plant habitats present. This sample typically covers a bank length of 5 m to 15 m .
3. Four net hauls from deeper bottom sediments along a canal length of approximately 10 m , elutriated on site to wash out the bulk of muds and fine sands. These should be taken at c .3 m intervals along the canal sampling length.

Two directly compatible field techniques can be employed to gather the four bottom sediment sample hauls from deeper areas, the choice depending on canal depth and accessibility:
(i) where canals are shallow enough to wade, bottom samples can be collected using a hand-net haul (c.3m length) taken perpendicular to the bank, (ii) where canals are too deep to use a hand net, bottom samples are collected using a Naturalist's dredge with a hand net sub-sample filling ca. one quarter of the pond net then taken from this dredged material. It is recommended that the bank and bottom samples are kept separate, since this makes the samples easier to sort in the laboratory.

### 4.5 Processing samples

Invertebrate sorting and identification methods follow the standard laboratory techniques used for processing invertebrate samples. Invertebrate samples are identified to family level for most groups and class level for oligochaetes.

Record findings in the columns on the field sheet as follows. If present and so included in ASPT calculation, record in the "ASPT" column, if a dragonfly or alderfly family also record in the "OM" column, or if a Coleoptera family in the "Cole." column.

### 4.6 Data processing and analysis

Invertebrate family data are used by PSYM to calculate four metrics:

- Average score per taxon (ASPT)
- Number of Ephemeroptera, Plecoptera and Trichoptera families (F_EPT)
- Number of beetle (Coleoptera) families (F_COL)
- Number of invertebrate families (INV_NFA)


### 4.7 Data interpretation and diagnosis

In analyses it was shown that ASPT and EPT scores both correlated strongly with a wide variety of water quality parameters, including heavy metals, suspended solids and chemical water quality (i.e. the overall chemical quality class based on suspended solids, BOD and ammonia concentrations). These metrics, however, showed few relationships with bank degradation variables.

In contrast, invertebrate family richness, and particularly beetle, bug and snail richness, showed strong relationships with bank structure and boat traffic, but very few relationships with water quality attributes.

These differences in degradation sensitivity make it possible to assess both water quality and bank effects separately. Thus where the main aim of canal assessments is to investigate water quality, then metrics based on ASPT and EPT taxa will be most effective. If boat traffic and hard bank structure effects are of concern, then parameters based on taxon richness or bug and beetle species or family richness can be combined into the final integrity index, i.e.:
A. Canal water quality assessment $=\mathrm{ASPT}+\mathrm{EPT}$.
B. Canal bank quality assessment $=$ No. Coleoptera families + No. invertebrate families.

Total canal ecological quality $=\mathrm{A}+\mathrm{B}$.

## Calculating the canal metrics from taxon lists

## 1. Average score per taxon (ASPT)

ASPT is calculated by summing the BMWP scores for all taxa present at the site and dividing by the total number of BMWP taxa present.

## 2. Number of Ephemeroptera, Plecoptera and Trichoptera families (F_EPT)

The sum of the number of Ephemeroptera, Plecoptera and Trichoptera families recorded in the sample.

## 3. Number of Coleoptera families ( $F_{-}$COL)

This metric is simply the sum of the number of Coleoptera families present at the site.

## 4. Number of invertebrate families (INV_NFA)

The number of all invertebrate taxa recorded on the survey form.

## 5. References and additional reading

Maitland, P.S. (1977). A coded checklist of animals occurring in fresh water in the British Isles. Institute of Terrestrial Ecology, Edinburgh.
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Pond Action (1998). A guide to the methods of the National Pond Survey. Pond Action, Oxford.
Preston, C.D., D.A. Pearman and T.D. Dines (Eds) (2002) New Atlas of the British and Irish Flora, Oxford University Press, Oxford.
Wright, J.F. (1995). Development and use of a system for predicting the macroinvertebrate fauna in flowing waters. Australian Journal of Ecology, 20, 181-197.
Wright, J.F., D. Moss, P.D. Armitage and M.T. Furse (1984). A preliminary classification of running-water sites in Great Britain based on macro-invertebrate species and the prediction of community type using environmental data. Freshwater Biology, 14, 221-256.

More detailed information describing the PSYM methodology is given in the following reports:
Williams, P., J. Biggs, L. Dodds, M. Whitfield, A. Corfield and G. Fox (1996). Biological techniques of still water quality assessment. Phase 1 Scoping Study. Environment Agency R\&D Technical Report E7. Environment Agency, Bristol.
Williams, P., J. Biggs, M. Whitfield, A. Corfield, G. Fox and K. Adare (1998). Biological techniques of still water quality assessment. Phase 2. Method development. Environment Agency R\&D Technical Report E56. Environment Agency, Bristol.
Biggs, J., P. Williams, M. Whitfield, G. Fox and P. Nicolet (2000). Biological techniques of still water quality assessment. Phase 3. Method development. Environment Agency R\&D Technical Report E110. Environment Agency, Bristol.

Pond PSYM Fieldsheet
Site and sample details

| Site name | Code No. | Grid ref. ( ) |
| :---: | :---: | :---: |
| Location |  |  |
| Site access details |  |  |
| Survey date | Surveyor |  |
| Notes |  |  |
| Environmental data | pH | Sketch of pond |
| Altitude (m) |  |  |
| Shade: \% pond overhung | \% emergent plant cover |  |
| Inflow (absent $=0$, present $=1$ ) | Pond area ( $\mathrm{m}^{2}$ ) |  |
| \% of pond margin grazed |  |  |

Pond base: categorise into one of three groups: $1=0 \%-32 \%, \quad 2=33 \%-66 \%, \quad 3=67 \%-100 \%$
Clay/silt
Sand, gravel, cobbles $\qquad$ Bed rock
Peat Other

## MACROINVERTEBRATE LIST



| $\begin{array}{\|c\|} \hline \mathbf{R} \\ \mathbf{S} \\ \hline \end{array}$ | TRS | Emergent plants | $\begin{array}{\|l\|} \hline \mathbf{R} \\ \mathbf{S} \end{array}$ | TRS |  | RS | TRS |  | RS | TRS | Submerged plants |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | Achillea ptarmica | 1 |  | Epilobium hirsutum | 1 | 7.3 | Phragmites australis | 2 | 6.3 | Apium inundatum |
| 1 |  | Acorus calamus | 1 |  | Epilobium obscurum | 4 | 5.5 | Pilularia globulifera | 1 |  | Aponogeton distachyos |
| 1 |  | Agrostis canina | 1 |  | Epilobium palustre | 2 |  | Pinguicula lusitanica | 1 |  | Cabomba caroliniana |
| 1 | LP | Agrostis stolonifera | 1 |  | Epilobium parviflorum | 1 |  | Pinguicula vulgaris | 2 |  | Callitriche brutia |
| 32 |  | Alisma gramineum | 2 |  | Epilobium tetragonum | 1 |  | Potentilla erecta | 1 | 6.3 | Callitriche hamulata |
| 2 |  | Alisma lanceolatum | 2 |  | Epipactis palustris | 1 | 5.3 | Potentilla palustris | 2 | 8.5 | Callitriche hermaphroditica |
| 1 | 9 | Alisma plantago-aquatica | 1 | LP | Equisetum fluviatile | 1 |  | Pulicaria dysenterica | 2 |  | Callitriche obtusangula |
| 2 |  | Alopecurus aequalis | 1 |  | Equisetum palustre | 16 |  | Pulicaria vulgaris | 2 |  | Callitriche platycarpa |
| 4 |  | Alopecurus borealis | 1 |  | Erica tetralix | 1 |  | Ranunculus ficaria | 1 | 7.3 | Callitriche stagnalis |
| 1 |  | Alopecurus geniculatus | 1 | 2.5 | Eriophorum angustifolium | 1 | LP | Ranunculus flammula | 4 |  | Callitriche truncata |
| 2 |  | Anagallis tenella | 16 |  | Eriophorum gracile | 2 | 10 | Ranunculus hederaceus | 1 |  | C. stagnalis/platycarpa agg. |
| 2 |  | Andromeda polifolia | 2 |  | Eriophorum latifolium | 2* |  | Ranunculus lingua | 1 |  | C. hamulata/brutia agg. |
| 1 |  | Angelica archangelica | 1 |  | Eriophorum vaginatum | 2 |  | Ranunculus omiophyllus | 1 |  | Callitriche sp. (undet.) |
| 1 |  | Angelica sylvestris | 1 |  | Eupatorium cannabinum | 32 |  | Ranunculus ophioglossifolius | 2 | 10 | Ceratophyllum demersum |
| 2 |  | Apium graveolens | 1 |  | Filipendula ulmaria | 32 |  | Ranunculus reptans | 2 |  | Ceratophyllum submersum |
| 1 | 10 | Apium nodiflorum | 2 |  | Galium boreale | 1 | 10 | Ranunculus sceleratus | 2 | 7.3 | Chara sp. |
| 32 |  | Apium repens | 8 |  | Galium constrictum | 2 |  | Rhynchospora alba | 1 |  | Egeria densa |
| 2 |  | Baldellia ranunculoides | 1 |  | Galium palustre | 4 |  | Rhynchospora fusca | 4 | 7 | Elatine hexandra |
| 2 | 10 | Berula erecta | 2 |  | Galium uliginosum | 2 |  | Rorippa amphibia | 4 |  | Elatine hydropiper |
| 2 |  | Bidens cernua | 1 |  | Geum rivale | 8 |  | Rorripa islandica | 2 |  | Eleogiton fluitans |
| 1 |  | Bidens connata | 2 |  | Glyceria declinata | 2 | 10 | Rorippa microphylla | 1 |  | Elodea callitrichoides |
| 1 |  | Bidens frondosa | 1 | LP | Glyceria fluitans | 1 | 10 | Rorippa nasturtium-aquaticum | 1 | 7.3 | Elodea canadensis |
| 2 |  | Bidens tripartita | 1 | 10 | Glyceria maxima | 1 | 10 | Rorippa (undet.). | 1 | 10 | Elodea nuttallii |
| 2 |  | Blysmus compressus | 2 |  | Glyceria notata | 1 |  | Rorippa palustris | 8 |  | Eriocaulon aquaticum |
| 2 |  | Bolboschoenus maritimus | 1 |  | Gnaphalium uliginosum | 2 | 10 | Rumex hydrolapathum | 1 | 6.3 | Fontinalis antipyretica |
| 2* |  | Butomus umbellatus | 1 | LP | Hydrocotyle vulgaris | 2 |  | Rumex maritimus | 2 |  | Groenlandia densa |
| 2 |  | Calamagrostis canescens | 2 |  | Hypericum elodes | 2 |  | Rumex palustris | , | 7.7 | Hippuris vulgaris |
| 2 |  | Calamagrostis epigejos | 1 |  | Hypericum tetrapterum | 1 |  | Sagina procumbens | 2 |  | Hottonia palustris |
| 8 |  | Calamagrostis purpurea | 4 |  | Hypericum undulatum | 1 |  | Sagittaria rigida | 4 |  | Isoetes echinospora |
| 8 |  | Calamagrostis stricta | 2 |  | Impatiens capensis | 2 |  | Samolus valerandi | , | 5 | Isoetes lacustris |
| 16 |  | Calamogrostis scotica | 1 |  | Impatiens glandulifera | 2 | 7.7 | Schoenoplectus lacustris | 1 |  | Lagarosiphon major |
| 1 |  | Calla palustris | 4* |  | Impatiens noli-tangere | 32 |  | Schoenoplectus pungens | 2 | 6.7 | Littorella uniflora |
| 1 | 7 | Caltha palustris | 1 | LP | Iris pseudacorus | 2 |  | Schoenoplectus tabernaemontani | 2 | 5 | Lobelia dortmanna |
| 1 |  | Cardamine amara | 1 |  | Isolepis setacea | 32 |  | Schoenoplectus triqueter | 8 |  | Ludwigia palustris |
| 1 |  | Cardamine pratensis | 1 |  | Juncus acutiflorus | 16 |  | Schoenus ferrugineus | 1 | 6.7 | Myriophyllum alterniflorum |
| 2 |  | Carex acuta | 1 |  | Juncus articulatus | 2 |  | Schoenus nigricans | 1 |  | Myriophyllum aquaticum |
| 1 | 10 | Carex acutiformis | 1 |  | Juncus bufonius agg. | 16 |  | Scorzonera humilis | 2 | 9 | Myriophyllum spicatum |
| 4 |  | Carex appropinquata | 1 | 5.3 | Juncus bulbosus | 1 |  | Scrophularia auriculata | 4 |  | Myriophyllum verticillatum |
| 2 |  | Carex aquatilis | 2 |  | Juncus compressus | 1 |  | Scutellaria galericulata | 4 |  | Najas flexilis |
| 2 |  | Carex curta | 1 |  | Juncus conglomeratus | 1 |  | Senecio aquaticus | 2 | 6.7 | Nitella sp. |
| 2 |  | Carex diandra | 1 | $L P$ | Juncus effusus | 1 |  | Senecio fluviatilis | 2 |  | Oenanthe fluviatilis |
| 1 |  | Carex disticha | 2 |  | Juncus foliosus | 32 |  | Senecio paludosus | 16 |  | Potamogeton acutifolius |
| 1 |  | Carex echinata | 1 |  | Juncus inflexus | 4 |  | Sium latifolium | 2 | 5.5 | Potamogeton alpinus |
| 2 | 10 | Carex elata | 32 |  | Juncus pygmaeus | 1 | 10 | Solanum dulcamara | 2 | 7.3 | Potamogeton berchtoldii |
| 4 |  | Carex elongata | 2 |  | Juncus subnodulosus | 4 |  | Sonchus palustris | 4 |  | Potamogeton coloratus |
| 1 |  | Carex flacca | 4 |  | Lathyrus palustris | 1 | 8.5 | Sparganium erectum |  |  | Potamogeton compressus |
| 1 |  | Carex hostiana | 32 |  | Leersia oryzoides | 1 |  | Stachys palustris | 1 | 10 | Potamogeton crispus |
| 2 |  | Carex laevigata | 32 |  | Liparis loeselii | 2 |  | Stellaria palustris | 16 |  | Potamogeton epihydrus |
| 2 | 4 | Carex lasiocarpa | 1 |  | Lotus pedunculatus | 1 |  | Stellaria uliginosa | 4 | 10 | Potamogeton filiformis |
| 2 | 4 | Carex limosa | 1 |  | Luzula luzuloides | 1 |  | Symphytum officinale | 2 | 10 | Potamogeton friesii |
| 1 | 5 | Carex nigra | 2 |  | Luzula sylvatica | 16 |  | Teucrium scordium | , | 7 | Potamogeton gramineus |
| 1 |  | Carex oedocarpa | 1 |  | Lychnis flos-cuculi | 2 |  | Thalictrum flavum | 2 | 10 | Potamogeton lucens |
| 1 |  | Carex otrubae | 1 |  | Lycopus europaeus | 4 |  | Thelypteris palustris | 8 |  | Potamogeton nodosus |
| 1 |  | Carex panicea | 1 |  | Lysimachia nummularia | 2 |  | Tofieldia pusilla |  | 8 | Potamogeton obtusifolius |
| 2 | 10 | Carex paniculata | 1 |  | Lysimachia terrestris | 1 |  | Trichophorum cespitosum | 1 | 10 | Potamogeton pectinatus |
| 1 |  | Carex pendula | 4 |  | Lysimachia thyrsiflora | 1 |  | Triglochin palustre | 2 | 7.3 | Potamogeton perfoliatus |
| 2 | 10 | Carex pseudocyperus | 2 |  | Lysimachia vulgaris | 2 | 10 | Typha angustifolia | 2 | 8.5 | Potamogeton praelongus |
| 1 |  | Carex pulicaris | 16 |  | Lythrum hyssopifolium | 1 | 8.5 | Typha latifolia | 2 | 9 | Potamogeton pusillus |
| 1 | 10 | Carex riparia | 2 |  | Lythrum portula | 2 |  | Valeriana dioica | 8 |  | Potamogeton rutilus |
| 1 | 5.3 | Carex rostrata | 1 |  | Lythrum salicaria | 1 |  | Vallisneria spiralis | 4 | 10 | Potamogeton trichoides |
| 2 |  | Carex spicata | 1 | 7.3 | Mentha aquatica | 1 |  | Veronica anagallis-aquatica | 2 | 10 | Ranunculus aquatilis |
| 2 |  | Carex vesicaria | 16 |  | Mentha pulegium | 1 | 10 | Veronica beccabunga | 2 | 10 | Ranunculus baudotii |
| 1 |  | Carex viridula | 1 | 5.3 | Menyanthes trifoliata | 2 |  | Veronica catenata | 2 | 10 | Ranunculus circinatus |
| 16 |  | Carex vulpina | 1 |  | Mimulus guttatus | 1 | 5.5 | Veronica scutellata | 2 |  | Ranunculus fluitans |
| 1 |  | Carex sp. | 1 |  | Mimulus luteus | 1 |  | Veronica sp. (undet.) | 2 | 7 | Ranunculus peltatus |
| 2 |  | Catabrosa aquatica | 16 |  | Minuartia stricta | 1 |  | Viola palustris | 2 | 8.5 | Ranunculus penicillatus |
| 4 |  | Cicuta virosa | 1 |  | Molinia caerulea | 32 |  | Viola persicifolia | 2 | 8.5 | Ranunculus trichophyllus |
| 2 |  | Cirsium dissectum | 1 |  | Montia fontana | 1 |  | Unknown exotic | 16 |  | Ranunculus tripartitus |
| 1 |  | Cirsium palustre | 1 | 7.7 | Myosotis laxa |  |  |  | 1 |  | Ranunculus sp. (undet.) |
| 2 |  | Cladium mariscus | 1 | 9 | Myosotis scorpioides | Floating-leaved plants |  |  | 1 |  | Sagittaria latifolia |
| 1 |  | Conium maculatum | 1 |  | Myosotis secunda | 1 |  | Azolla filiculoides | 1 |  | Sagittaria sagittifolia |
| 1 |  | Crassula helmsii | 4 |  | Myosotis stolonifera | 2 |  | Hydrocharis morsus-ranae | 2 |  | Sagittaria subulata |
| 1 |  | Crepis paludosa | 1 |  | Myosotis sp (undet.). | 1 |  | Hydrocotyle ranunculoides | 2 | 4 | Sparganium angustifolium |
| 16 |  | Cyperus fuscus | 2 |  | Myosoton aquaticum | 2 |  | Lemna gibba | 1 | 10 | Sparganium emersum |
| 4* |  | Cyperus longus | 1 |  | Myrica gale | 1 | 9 | Lemna minor | 2 |  | Sparganium natans |
| 2 |  | Dactylorhiza sp (undet.) | 1 |  | Narthecium ossifragum | 1 |  | Lemna minuta | 1 | 2.5 | Sphagnum sp. |
| 32 |  | Damasonium alisma | 2 |  | Oenanthe aquatica | 1 | 10 | Lemna trisulca | 4* |  | Stratiotes aloides |
| 1 |  | Deschampsia cespitosa | 1 |  | Oenanthe crocata | 4 |  | Luronium natans | 2 | 4 | Subularia aquatica |
| 2 |  | Drosera anglica | 2 |  | Oenanthe fistulosa | 1 |  | Nuphar advena |  |  | Tolypella sp. |
| 1 |  | Drosera binata | 2 |  | Oenanthe lachenalii | 2 | 8.5 | Nuphar lutea | 2 |  | Utricularia australis |
| 1 |  | Drosera capensis | 2 |  | Oenanthe pimpinelloides | ${ }^{4}$ | 7 | Nuphar pumila | 2 | 4 | Utricularia intermedia |
| 2 |  | Drosera intermedia | 4 |  | Oenanthe silaifolia | $2^{*}$ | 6.7 | Nymphaea alba | 2 | 4 | Utricularia minor |
| 16 |  | Drosera rotundifolia | 2 |  | Osmunda regalis | 4* |  | Nymphaea sp. (exotic) | 2 | 5 | Utricularia vulgaris |
| 16 |  | Dryopteris cristata | 2 |  | Parnassia palustris | 4* |  | Nymphoides peltata | 1 |  | Vallisneria spiralis |
| 2 |  | Eleocharis acicularis | 1 |  | Pedicularis palustris | 1 | , | Persicaria amphibia | 2 | 10 | Zannichellia palustris |
| 8 |  | Eleocharis austriaca | 1 | 10 | Persicaria hydropiper | 1 | LP | Potamogeton natans |  |  |  |
| 2 |  | Eleocharis multicaulis | 1 |  | Persicaria maculosa | 1 | 3.7 | Potamogeton polygonifolius |  |  |  |
| 1 | $L P$ | Eleocharis palustris | 2 |  | Persicaria minor | 2 |  | Riccia fluitans |  |  | submerged species |
| 2 |  | Eleocharis quinqueflora | 4 |  | Persicaria mitis | 2 |  | Ricciocarpus natans |  |  | Number of uncommon species |
| 2 |  | Eleocharis uniglumis | 1 |  | Petasites hybridus | 2 |  | Spirodela polyrhiza |  |  | (with a rarity score of 2 or more) |
| 2 |  | Epilobium alsinifolium | 1 |  | Petasites japonicus | 4 |  | Wolffia arrhiza |  |  | Trophic Ranking Score |
| 2 |  | Epilobium anagallidifolium | 4 |  | Peucedanum palustre | * $=$ uncommon species often introduced to sites (see Preston et al. 2002 for details), if so score species as 1 . |  |  | $L P=$ species exhibiting little nutrient preference |  |  |
| 1 |  | Epilobium brunnescens | 1 | 8.5 | Phalaris arundinacea |  |  |  |  |  |  |
| 1 |  | Epilobium ciliatum |  |  |  |  |  |  |  |  |  |

## Canal PSYM Fieldsheet

Site and sample details

| Site name |  | Code no. <br> Northing | $\underset{\text { Recording format: (SU)3450 }}{6780 \text { or (41)3450 } 6780}$ |  |
| :---: | :---: | :---: | :---: | :---: |
| Easting |  |  |  |  |
| Canal |  | Location |  |  |
| Survey date | Surveyors |  | bank sampled |  |

## Environmental data



Macroinvertebrate list




Valvatidae
Hydrobiidae (Bithyniidae
Lymnaeidae
Physidae
Planorbidae
Sphaeriidae
Glossiphoniidae
Hirudinidae
Erpobdellidae
Asellidae


| Group 8 taxa (BMWP:2) |
| :--- |
| Chironomidae |

Group 9 taxa (BMWP:1) Oligochaeta

No. of taxa
TOTAL NO. OF TAXA

TOTAL BMWP SCORE

## ASPT

NO. OF EPT TAXA

NO. OF COLEOPT.


[^0]:    ${ }^{1}$ Metrics are variables such as species richness or rarity which can be used to help identify how damaged a waterbody's community is. They have been shown to have a strong monotonic relationship with degradation.
    ${ }^{2}$ RIVPACS. The River InVertebrate Prediction And Classification System, developed by the Institute of Freshwater Ecology and Environment Agency (Wright et al. 1984, Wright 1995).

[^1]:    ${ }^{3}$ Waterbodies between $1 \mathrm{~m}^{2}$ and 2 ha in area which usually retain water throughout the year (Collinson et al,. 1994). Includes both man-made and natural waterbodies.

[^2]:    ${ }^{4}$ Note that there is only one family of Megaloptera in the UK (the Sialidae) and that the metric F_OM is concerned with the combined total of Odonata and Megaloptera, not the occurrence of the family Megaloptera alone.

[^3]:    ${ }^{5}$ The rarity status values for Scarce and RDB species are based on existing definitions derived from the Red Data Books and other authorities. The definition of 'local' has been used to define species which are not uniformly common and widespread in Britain: with plants this refers specifically to species recorded from between 101 and $70010 \times 10 \mathrm{~km}$ squares (approximately $25 \%$ of all 10 km in England, Wales and Scotland).
    ${ }^{6}$ BMWP (Biological Monitoring Working Party) scores assigned to taxa defined by Maitland (1977), so each is allocated a value from 1 to 10 depending on its known tolerance to organic pollution, a higher score indicates lower tolerance.

[^4]:    ${ }^{7}$ Ideally PSYM should also include a plant-based assessment, however this has not yet been developed. In canals, diatoms have been identified as the most suitable plant assemblage for assessing quality, since macrophytes often occur in very low abundance where water is at all turbid and banks are reinforced.

